For Review Only, AAQR



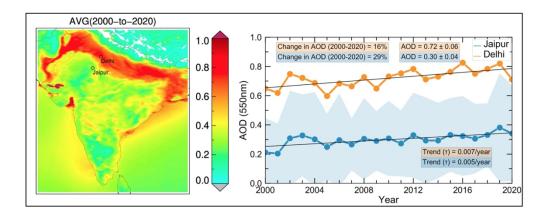
Temporal and Spatial Variability in Particulate Matter Pollution in The Pink City during the 2020 Diwali Festival

Journal:	Aerosol and Air Quality Research
Manuscript ID	Draft
Manuscript Type:	Special issue: Air Pollution and its Impact in South and Southeast Asia
Date Submitted by the Author:	n/a
Complete List of Authors:	Yadav, Piyush; Delhi Technological University, Electronics and Communication Engineering; RESOLVE NGO Gupta, Pawan; Universities Space Research Association, STI; NASA Marshall Space Flight Center,
Keywords:	Aerosol optical depth (AOD), Spatio-temporal patterns, Mobile PM2.5 monitoring, Low-cost sensor, Diwali

SCHOLARONE[™] Manuscripts

Highlights

- Studied $PM_{2.5}$ and long-term (2000–2020) aerosol optical depth of Jaipur city.
- Observed a 29% increase in aerosol optical depth of Jaipur, but was significantly lower than Delhi.
- Studied the temporal and spatial variability in PM_{2.5} of Jaipur during the 2020 Diwali festival.
- Observed significantly lower PM_{2.5} in Jaipur during 2020 Diwali compared to previous years due to favourable weather conditions and firecracker ban.
- Increase in aerosol optical depth and PM_{2.5} due to crop residue burning was also observed in Jaipur during Diwali week.
- With mobile air quality monitoring, location of a pollution hotspot was also identified.



312x124mm (144 x 144 DPI)

2 3

4 5 6

7

8 9

10 11

12

13 14

16

Temporal and Spatial Variability in Particulate Matter Pollution in The Pink City during the 2020 Diwali Festival

Piyush Yadav^{1,2*}, Pawan Gupta^{3, 4}

¹ Delhi Technological University, Delhi, 110042, India
 ² RESOLVE, Washington DC, 20037, USA
 ³ Universities Space Research Association (USRA), Huntsville, AL, 35758, USA
 ⁴ NASA Marshall Space Flight Center, Huntsville, AL, 35758, USA

* Corresponding author. Tel: 91-8890170383 E-mail address: pyadav@resolve.ngo

15 Abstract

17 The temporal and spatial variability of PM_{2.5} in Jaipur during the Diwali festival in the year 2020 is analysed and compared with previous years. Diwali festival is celebrated widely across India 18 every year by bursting firecrackers, which significantly deteriorates the short-term air quality. For 19 20 the temporal and spatial analysis, we used data from satellite, regulatory ground monitors, and mobile monitoring using a low-cost sensor. The climatological mean Aerosol Optical Depth (AOD) 21 22 derived from the Moderate Resolution Imaging Spectroradiometer (MODIS) Aqua and Terra sensor over Jaipur for the period 2000-2020 was 0.3 ± 0.04 . The linear trend analysis over Jaipur 23 shows an increase in annual mean AOD at a rate of 0.005/year, with a significant increase in AOD 24 levels (28.9%) in the past two decades. The increase in AOD is mainly due to the increase in dust 25 particles from the Thar desert, enhanced urban activities, and post-monsoon crop-fires. During the 26 Diwali festival in the previous years (2017-19), hourly PM_{2.5} concentrations were exceptionally 27 high during the night-time, whereas due to the firecracker ban and under favourable weather 28 29 conditions in the year 2020, no major spike in hourly PM25 concentrations was observed. The daily mean PM_{2.5} concentrations in the previous years (2017-19) were substantially higher Post-Diwali 30 than Pre-Diwali days, whereas the opposite trend was observed during Diwali 2020 due to 31 relatively higher wind speeds. In 2020, AOD levels showed a significant spike on 1-day Post-32 Diwali, whereas PM_{2.5} concentrations from surface measurements showed a drop. Satellite 33 observations show a layer of smoke over Jaipur transported from crop-fires in Northern India, that 34 lead to a significant rise in both the AOD and PM2.5 concentration on the 4th day Pre-Diwali. PM2.5 35 concentrations observed from mobile monitoring were higher Pre-Diwali days than Post-Diwali 36 and did not show much variation during the evening and night-time of Diwali day. 37

38 39

Keywords: Aerosol optical depth (AOD), Spatio-temporal patterns, Mobile PM_{2.5} monitoring,
 Low-cost sensor, Diwali

- 42
- 43
- 44
- 45

47 1 INTRODUCTION

48

The promising economic opportunities offered by urbanization, also bring a health penalty due to the rise in air pollution, especially for low-income countries (Wang, 2018). According to WHO, 9 out of 10 people in the world breathe poor air that exceeds the WHO guideline limit of air pollutants (WHO Air Pollution). This bad air quality has drastic effects on the health of both humans and climate, and every year the combination of ambient and household air pollution causes about seven million premature deaths worldwide (WHO Air Pollution), making the population in countries like India extremely vulnerable due to the severity of air pollution in the region.

According to the State of Global Air (2020), the population-weighted annual average PM_{2.5} 56 concentration of India increased by 6.5 μ g/m³ between 2010-19, with the annual average PM_{2.5} 57 exposure observed as 83.2 µg/m3 for 2019, which was the highest in the world. It was found that 58 59 in 2019, air pollution lead to 1.67 million deaths in India, which is 17.8% of the total deaths in the country, leading to an economic loss of \$36.8 billion in 2019 alone (Pandey et al., 2020). The State 60 61 of Global Air (2019) states that air pollution is one of the top three killers in India and it is estimated 62 that more than half of the world's pollution-related deaths occur in India and China collectively. 63 China has taken strict actions to improve the air pollution situation, by reforms such as banning new coal power plants in polluted areas, introducing massive afforestation, and reforestation 64 65 (Leung, 2021). The introduction of The Air Pollution Action Plan, in 2013 and the Three-year Action Plan for Winning the Blue Sky War, in 2018 have helped China to significantly improve its 66 67 air quality, whereas India is still in its early days in the fight against air pollution with the 68 introduction of National Clean Air Program in 2019 and an ambitious push towards clean energy 69 with a target to achieve 450 GW renewable energy capacity by 2030, of which 100 GW is installed

For Review Only, AAQR

70	till now (PIB). But activities like stubble burning and bursting fireworks during the Diwali festival
71	in India hinder India's fight against air pollution. (Chattopadhyay, 2014).
72	Air pollution has gained attention in recent years from all parts of society in northern India with
73	the pollution levels staying in the severe category in many cities (AQLI, 2021). For the 5,000 cities
74	and towns in India, it was found that around 4000 continuous ambient air quality monitoring
75	stations (CAAQMS) would be required to efficiently monitor air quality variations spatially and
76	temporally but as of September 2019, just 200 CAAQMS are operational in India (India – Ambient
77	Air Monitoring Data).
78	Due to rapid urbanization and population growth, Jaipur city (also known as the Pink City)
79	ranked as the 14th most polluted city in the world in 2016 (WHO Global Urban Ambient Air
80	Pollution Database, 2016) and was found to be the 22 nd most environmentally vulnerable city of
81	the world in 2021 (Environmental Risk Outlook, 2021).
82	Several past studies reported air pollution issues in the city but gaps in various aspects of
83	spatiotemporal variability remain. The majority of the previous studies used ground-based point
84	measurements to address air pollution issues and focused on a certain time period only. There is
85	still a lack of understanding on spatial and temporal scales of pollutions in the city, specifically
86	during the major festival such as Diwali. We attempt to address and fill in gaps on spatiotemporal
87	variability of PM (or aerosols) in Jaipur during Diwali while using 2 decades of satellite
88	observations of aerosols, multiple regulatory grade air quality monitors, and mobile monitoring
89	using a low-cost sensor. Section 2 provides details on the study area, data and method applied;
90	section 3 presents results on satellite-derived long-term trends, pollution during Diwali and role of
91	meteorology, and spatiotemporal analysis of mobile monitoring; section 4 discusses the results and
92	summary of major findings are reported in section 5.

94 2 METHODS

95

96 2.1 Study Area

In this study, air quality data was analysed for Jaipur, the capital city of the northwestern state
of Rajasthan in India. Jaipur has a population of 3.1 million and is the 10th most populous city of
India according to the census of 2011 (Census India, 2011).

Jaipur (26.91 °N, 75.78 °E) lies in the foothills of the Aravalli mountain range, surrounded by hillocks on its northern and eastern sides. It has a vast stretch of plains and desert on its western and southern sides. Being located downwind of the Thar desert, dust storms are quite frequent in the region. Jaipur is located about 268 km southwest of the national capital Delhi and witnesses extreme temperatures both in summer as well as in winters and low to moderate RH.

105 We analyzed the air quality data of Jaipur during the Diwali festival, also known as the festival 106 of lights. The Diwali festival typically falls in the month of October or November (observed 107 according to the moon calendar) every year (Table S1). During the week-long celebration of 108 Diwali, among other practices, the use of firecrackers is common across the country. The use of firecrackers usually results in poor air quality for a short period of time and is well documented 109 110 and reported in previous studies in different parts of the country (Attri et al., 2001; Kulshrestha et 111 al., 2004; Ganguly, 2009; Deka and Hoque, 2014; Saha et al., 2014; Ganguly, 2015). This is also the time of year when large-scale crop residual burning occurs in several adjunct states (Mittal et 112 113 al., 2009) and transported smoke can significantly increase the pollution concentration in downwind areas. 114

According to the study by <u>Pozzer et al. (2020)</u> on the coronavirus pandemic (COVID-19) that emerged in 2019, air pollution was found to be an important cofactor in increasing the risk of mortality from COVID-19 and estimated to have resulted in 15% of worldwide deaths from 118 COVID-19. PM_{2.5} data from ground stations were used for the year 2014 and found that PM_{2.5} concentrations of Jaipur city were significantly higher during the Diwali day (186 µg/m³) compared 119 120 to normal days before $(71\mu g/m^3)$ and after Diwali $(40 \mu g/m^3)$ and the air pollution condition in 121 Jaipur were found to be poor (201<AQI<300) or very poor (301<AQI<400) for a small fraction of 122 days during Diwali (Ganguly et al., 2019), with the AQI mentioned here corresponding to the 123 maximum value of the individual sub-indices of eight pollutants (PM₁₀, PM_{2.5}, NO₂, SO₂, CO, O₃, 124 NH₃, and Pb). So, in order to prevent the rise in mortality rates from Covid-19 due to the potential spike in air pollution during the Diwali festival in the year 2020, the Government of Rajasthan 125 126 imposed a statewide ban on the sale and bursting firecrackers to control the air pollution and 127 subsequently Covid-19 mortalities (The Times of India, Twitter).

128

129 **2.2 Surface Measurements**

130 PM_{2.5} is routinely measured from a reference-grade monitor located at three locations (Police 131 Commissionerate, Shastri Nagar, Adarsh Nagar) in the city (Fig. S9) and operated by Rajasthan 132 State Pollution Control Board (RSPCB). The measurement frequency of the monitoring station is 133 15 minutes. Reference grade monitor (i.e. BAM) data for the period of 2017-2020 is downloaded 134 from the Central Pollution Control Board (CPCB) database (CPCB, 2020) for this study. In the past studies, data obtained from CPCB monitoring stations were found to have measurement errors of 135 136 around less than 5% (Sarkawt et al., 2020; CPCB, 2011). To quality control the data, we only used hourly and daily mean concentrations from the CPCB database and eliminated PM_{2.5} 137 concentrations less than 2µg/m³. Reference grade monitors are also equipped with weather 138 139 monitors and provide data on wind, temperature, rain, and RH at the same temporal frequency.

140

141 2.3 Satellite Measurements

142 The MODIS sensor onboard NASA's Terra and Aqua satellite has been providing AOD data 143 since 2000 and 2003, respectively. The Terra and Aqua satellites have overpass times of 10:30 am 144 and 1:30 pm local solar time respectively. The MODIS collection 6.1, Level 2 AOD at 550nm with a spatial resolution of 10x10 km² of highest quality were used in the study. For analysing the annual 145 146 AOD trends in Jaipur, the daily AOD data derived from the two MODIS sensors and processed and 147 gridded following the level 3 method explained in our earlier work (i.e. Gupta et al., 2020). For analysing the day-to-day evolution of AOD during the Diwali festival, AOD data for ± 14 days of 148 149 Diwali for the study period (2005-2020) were retrieved from the high-resolution level 3 dataset 150 (Gupta et al., 2020). Similar to the MODIS sensor, the Visible Infrared Imaging Radiometer Suite 151 (VIIRS) sensor onboard NASA's Suomi-NPP satellite has been providing AOD data since 2011. 152 For analysing the aerosol transport to Jaipur city, the Corrected Reflectance (True Color) and Deep Blue AOD maps derived from the VIIRS sensor are downloaded from 10th November and 153 27th November 2020 for a 2 day period during Diwali using https://worldview.earthdata.nasa.gov/. 154

155

156 2.4 Mobile Measurements

The mobile air quality monitoring experiment in the city was performed by integrating low-cost Purple Air Sensor (PAS) and U-Blox Global Positioning System (GPS) module on the roof of a car (Fig. S1). The measurement height was 1.9 meters above the ground. The system design (Fig. S1) for the low-cost mobile air quality monitoring network was as follows: a) the measurement frequencies of the PAS and GPS module was 1 second (Yadav, 2020); b) the data from PAS and GPS got stored in a CSV file on the RaspberryPi (RP); c) the GPS and RP were powered by a 3s (12.6v) LiPo battery and the PAS was powered by a 5v power bank.

164 The PAS uses a PMS5003 Plantower sensor. It uses a factory calibration-based algorithm to 165 calculate the PM₁, PM_{2.5}, and PM₁₀ mass concentration in μ g/m³ (<u>PurpleAir</u>). The PAS currently

For Review Only, AAQR

outputs 2-minute average data, which makes the sensor not ideal for mobile sensing applications.
We modify the sensor outputs frequency and accessed the real-time instantaneous data from PAS
over wifi by running the script (<u>Yadav, 2020</u>) on RP. This new setup allowed us to gather air quality
measurements as well as geolocation at higher temporal resolution.

We collected the data from the mobile monitoring experiment for a period of 2 days Pre-Diwali (11th November and 13th November), 2 days Post-Diwali (15th November & 16th November), and during Diwali day (14th November: evening and night) to analyze the spatiotemporal PM_{2.5} variations in the city. The PAS data were only collected during the 2020 Diwali festival and to capture spatial patterns in air pollution in different parts of the city, data were collected on multiple routes (Fig. 5). In addition, to verify the output of PAS, several measurements were taken next to a reference-grade monitor (Shastri Nagar Station) and results are presented in Fig. S8.

177

178 2.5 Method

Air quality indicators (AOD and $PM_{2.5}$) and meteorological parameters (Wind, Temperature, RH) were used to understand the spatial and temporal patterns. In absence of long-term ground $PM_{2.5}$ data, we used satellite AODs from MODIS for the past 20 years to analyse annual trends whereas $PM_{2.5}$ data was used to see short-term (2017-2020) trends.

To quantify the temporal patterns of air pollution during the Diwali festival in Jaipur, we performed the following analysis: a) hour-to-hour evolution of $PM_{2.5}$ concentration for \pm 3 days of Diwali from 2017-2020; b) day-to-day change in $PM_{2.5}$ concentration for \pm 14 days of Diwali from 2017-2020; c) day-to-day evolution of AOD for \pm 14 days of Diwali from 2005-2020. For analysing the temporal patterns in $PM_{2.5}$ concentrations during the 2020 Diwali festival, we used 2017-2019 mean as the baseline condition, as this was the only data available for previous years from the CPCB database. While for the temporal patterns in AOD, we used 2005-2019 mean as the baseline. We also used meteorological parameters to understand their role in temporal patterns of $PM_{2.5}$ and AODs. The transport patterns of pollution were qualitatively evaluated using VIIRS true-color images and AOD maps. To quantify the spatiotemporal variations of air pollution during the Diwali festival, we performed mobile monitoring with a low-cost sensor using the time frame of Pre, Post, and during Diwali day on 3 different paths in the city.

195

196 **3 RESULTS**

197

198 **3.1 Trend in AOD and PM_{2.5}**

AOD is a measure of the amount of light extinction (by scattering or absorption) by aerosols in 199 200 the atmospheric column. AOD is measured at specific wavelengths, with 550nm being the common 201 reference wavelength used by satellite AOD products. In the past, many studies (Christopher and Gupta, 2020; Gupta et al., 2006; Liu et al., 2009; Sathe et al., 2019; Donkelaar et al., 2019; Yang 202 203 et al., 2019) have shown the relationship between AOD and surface PM_{2.5} and found that it is not 204 always linear. The estimation of PM2.5 from AOD required complex statistical modeling with 205 additional ancillary data sets. Thus, we have not derived PM_{2.5} from AOD in this study, rather have 206 used AOD as an independent parameter to monitor air pollution variation. The 20-Year mean AOD 207 distribution over India (Fig. 1(a)) for the period 2000-2020 shows highly elevated AOD levels in 208 the Indo-Gangetic Plain (IGP). The IGP is a densely populated region and is home to several 209 industries and power plants, which contribute to pollutions. This severe air pollution scenario and 210 its cause in the region has been well documented in the past (Kumar et al., 2018; Devi et al., 2020; 211 Ojha et al., 2020). The marked circles on the map in Fig. 1(a) show the location of Jaipur and Delhi. 212 Compared to Delhi (AOD=0.72), the 20-Year mean AOD levels observed over Jaipur (AOD=0.3)

For Review Only, AAQR

are lower by a factor of more than 2. Using level 2 and level 3 monthly aerosol products obtained
from MODIS, MIRS and Aerosol Robotic Network (AERONET) sensor, (Mangla et al., 2020)
showed that for the period of 2010-2017, AOD values in Jaipur were lower than other regions in
IGP.

217 Fig. 1(b) shows the time-series of annual mean AODs for the past two decades. The marker represents the annual mean AOD and the shaded area represents one standard deviation (SD). The 218 219 AOD over both Jaipur and Delhi showed a consistent and increasing trend, with trends being 0.005 220 and 0.007 per year for the two cities respectively. The AOD values increased to (0.34 ± 0.33) from 221 (0.21 ± 0.23) for Jaipur over a period of 20 years, which is equivalent to about a 29% increase. Jaipur being an arid region and its proximity to Thar desert, Ramachandran et al. (2012) showed 222 223 that the increase in wind speed observed over Jaipur from 2000-2009 lead to an increase in soilderived dust particles, which subsequently resulted in an increasing annual AOD trend (43.2%) 224 225 over Jaipur. While the increasing trend in AOD over Delhi was attributed to an increase in the 226 amount of aerosols from fossil fuel and biomass burning (Ramachandran et al., 2012). By 227 converting satellite-derived AOD to surface PM_{2.5}, <u>Dey et al. (2020)</u> found that between 2000-19, 228 the Thar desert showed a decreasing trend in PM_{2.5} while Jaipur and Delhi showed an increasing 229 trend due to the post-monsoon burning activities (Chowdhury et al., 2019). Whereas Mangla et al. (2020) reported high AOD over Jaipur during the months of April to June for the period 2010-2017 230 231 due to dust activities from the Thar Desert, Arabian Peninsula, and Gulf regions (Verma et al., 232 <u>2013)</u>.

Fig. 1(c) shows the annual box-plot of $PM_{2.5}$ concentrations over Jaipur and Delhi for the period 2017-2020. In Fig. 1(c), the box represents 25th and 75th percentiles, whisker represents the 235 minimum and the maximum, horizontal line represents median and the circle represents mean. The 236 annual mean $PM_{2.5}$ observed in Jaipur was 78 ± 53.2, 75.6 ± 32.9, 60.9 ± 27.7, and 53.4 ± 24.4

237	μ g/m ³ between the period 2017-2020 respectively. Compared to Delhi (102.6 ± 93.5), the 4-Year
238	mean $PM_{2.5}$ observed over Jaipur (67 ± 34.6) is lower by a factor of 1.5.
239	Although the annual mean $PM_{2.5}$ over Jaipur has been relatively lower than in Delhi, it is still
240	remarkably higher than the annual mean WHO air quality guideline (10 μ g/m ³).
241	
242	3.2 Air Pollution during Diwali as viewed from Surface and Satellite
243	The use of firecrackers is common to celebrate Diwali in India and it often generates harmful
244	pollutants. Here, we analysed the hourly evolution of $PM_{2.5}$ from surface measurements in Fig. 2(a)
245	for \pm 3 days of Diwali in 2020 and the previous years (2017-19), to assess the impact of firecrackers
246	and meteorology on air pollution in the city.

 \pm 14 days of Diwali in 2020 and the previous years (2017-19), while Fig. 2(c) gives the long-term perspective of air pollution during Diwali using the mean AOD data from the two MODIS sensors, showing the day-to-day evolution in AOD for \pm 14 days of Diwali in 2020 and the previous years (2005-19).

252

3.2.1 Diurnal and Day-to-Day Variability

Typically during Diwali, the peak firecracker bursting activities start late in the evening and last till around midnight and it extends about ± 3 days around Diwali day. Hourly PM_{2.5} for the previous years (2017-19) show an increase in value starting around 7 pm, peaks around midnight, and remains high until the next morning. The peak hourly PM_{2.5} during Diwali night was reported as 834, 622, 881, and 149 µg/m³ for 2017, 2018, 2019, and 2020 respectively. This pattern of rise in PM_{2.5} concentrations at night-time can be observed on Diwali day and for 2 days following Diwali Page 13 of 52

For Review Only, AAQR

in the previous years (2017-19). But in 2020 (red line on in Fig. 2(a)), no rise in hourly PM_{2.5} values
was observed.

This periodic rise in $PM_{2.5}$ concentrations during Diwali week in the previous years (2017-19) is associated with firecrackers, but the firecracker ban in 2020 seemed to have stopped this pattern during the Diwali festival.

The surface $PM_{2.5}$ analysis is extended in Fig. 2(b) using daily mean values for \pm 15 days of Diwali to understand the impact on daily mean values from the use of firecrackers. Similar to hourly data, daily mean data consistently shows lower values in 2020 compared to previous years. Fig. 2(b) shows that overall $PM_{2.5}$ values in the pre-Diwali period were consistently lower in 2020 compared to other years, whereas they were higher in the post-Diwali period.

The daily mean (table 1) for Diwali for 2017-2019 varies between 114 ± 136 and 245 ± 231 whereas remains low in 2020 (106 ± 37). Table 1 shows the daily mean PM_{2.5} \pm SD value for ± 3 days of Diwali in 2020 and the previous years (2017-19). From Table 1, we can observe the drop in daily mean PM_{2.5} concentrations for days following Diwali in 2020, with daily mean PM_{2.5} being 106 ± 37 on Diwali day and 79 ± 15 , 63 ± 15 , $56 \pm 18 \,\mu$ g/m³ for the 3-day period following Diwali respectively. The daily mean PM_{2.5} showed a decreasing trend in 2020 after Diwali, whereas it showed an increasing trend in previous years (2017-19).

277

278 3.2.2 Impact of meteorology

Meteorological factors such as wind speed, temperature, and RH have a significant influence
on air pollution patterns (<u>Cichowicz et al., 2020</u>; <u>Liu et al., 2020</u>; <u>Chen et al., 2020</u>; <u>Jayamurugan</u>
<u>et al., 2013</u>; <u>Dey et al., 2017</u>). Fig. S2 shows the hourly wind speed data for ±3 days of Diwali in
2020 and the previous years (2017-19). Typically winds are calm during early winter in the City.
Hourly wind speed observed at midnight (12 am) on Diwali day were 0.61, 1.02, 0.05, and 1 m/s

for the period 2017-2020 respectively. Despite the similarity in wind speeds at 12 am on Diwali

284

285 night in 2018 and 2020, the PM_{2.5} concentrations between 2018 and 2020 differed significantly 286 and were observed as 622 and 149 μ g/m³ respectively. This clearly shows a factor other than just 287 higher wind contributing to lower PM_{2.5} concentrations in 2020. 288 From Fig. S2, we can also observe a periodic rise in the hourly wind speed for the years 2018-289 2020 during the Diwali week that starts at around 7 am in the morning and lasts till around 7 pm 290 in the evening, with wind speeds peaking at around 12 pm. Due to the elevated wind speeds 291 during the afternoon time (Fig. S2), from Fig. 2a we can observe a periodic dip in the hourly 292 PM_{2.5} concentrations that starts at 7 am in the morning and lasts till around 7 pm in the evening, 293 with the lowest concentrations observed at around 13 pm. This increasing trend in the hourly 294 wind speed between 7 am to 7 pm, leads to a decreasing trend in the hourly PM_{2.5} concentrations between 7 am to 7 pm during the \pm 3 days of the Diwali week. 295 296 Fig. S3 shows the daily box plot of wind speed for \pm 3 days of Diwali in 2020 and the 297 previous years (2017-19). Compared to previous years (2017-19), the daily mean wind speed 298 shows a significant rise Post-Diwali in 2020, thus leading to the drop in daily mean PM_{2.5} 299 concentrations in 2020, as observed in Fig. 2b and Table 1. Fig. S6 also shows an increasing 300 trend in the daily mean values of both wind speed and RH Post-Diwali day in 2020, thus leading 301 to a drop in $PM_{2.5}$ levels Post-Diwali. 302 Fig. S4 and Fig. S5 show the scatter plot of wind speed and RH with PM_{2.5} respectively using 303 the dataset of ± 3 days of Diwali for 2020 and the previous years (2017-19). Fig. S4 shows the 304 expected inverse relationship between wind speed and PM_{2.5} in the previous years (2017-19), 305 with values going beyond 600 µg/m³ at low wind speeds. Whereas in 2020, the PM_{2.5} 306 concentrations were always below 200 µg/m³ even at low wind speeds. From Fig. S5, we can

> 12 Editorial Office : Web page: http://aaqr.org/ Tel: 886-3-5731880 Fax: 886-3-5727835

307 observe that RH was significantly higher in 2020 than in previous years, thus this could one of
308 the reasons for lower PM_{2.5} concentrations observed in 2020.

309

310 **3.3 Spatial distribution of air pollution during Diwali 2020**

311

312 3.3.1 Citywide monitoring stations

Often air quality variability within a city depends on many factors, such as meteorology (Cichowicz et al., 2020; Liu et al., 2020; Chen et al., 2020; Jayamurugan et al., 2013; Dey et al., 2017), vehicular emissions (Sindhwani and Goyal, 2014; Guttikunda et al., 2014) and traffic density and patterns (Thaker and Gokhale, 2016), and short and long-range transport. To analyse the spatial variation of air pollution in Jaipur city during Diwali 2020, we evaluated the PM_{2.5} and wind speed data of Jaipur's 3 monitoring stations.

319 Fig. 3(a) shows the daily mean $PM_{2.5}$ concentrations for ±14 days of Diwali 2020 for the 3 320 monitoring stations of Jaipur. The 3 monitoring stations showed a high correlation in daily mean 321 PM_{2.5} concentrations, thus showing the less spatial variability exhibited by PM_{2.5} in Jaipur during 322 Diwali week in 2020. Due to the increasing trend in wind speeds (Fig. 3(b)) from 1-day Pre-Diwali 323 to 1-day Post-Diwali, there was a decreasing trend in PM_{2.5} (Fig. 3(a)) between 1-day Pre-Diwali 324 to 2nd day Post-Diwali. The wind speed observed at the Adarsh Nagar monitoring station in Fig. 325 3(b) was found to be significantly higher than the other two stations. But these higher wind speeds 326 observed at Adarsh Nagar station didn't seem to have any visible impact on PM_{2.5} concentrations, 327 as all the stations seem to have similar concentrations, despite a significant difference in wind 328 speeds.

Fig. S7 shows the hourly PM_{2.5} and wind speed data of the city's 3 monitoring stations for the day period (1-day Pre-Diwali to 3rd day Post-Diwali) during Diwali. From Fig. S7, it can be seen that at 12 am on Diwali night, Adarsh Nagar and Shastri Nagar monitoring stations had higher wind speed (1.71 m/s and 1.85 m/s respectively) than the Police Commissionerate monitoring station (1 m/s), and due to the higher wind speeds, lower $PM_{2.5}$ concentrations (93 µg/m³ and 90 µg/m³ respectively) than Police Commissionerate station (149 µg/m³).

335

336 **3.3.2 Impact of aerosol transport**

To track and monitor the long-range transport of aerosols around the world, satellite data has 337 338 acted as a very reliable source. Some common types of aerosol transport are from biomass burning 339 (Gupta et al., 2007; Huff et al., 2017; Cusworth et al., 2018; Wang and Christopher, 2006), dust 340 storms (Yu et al., 2019; Naeger et al., 2016), and volcanic eruption (Flower and Kahn, 2020), which 341 occur at various spatiotemporal scales. Fig. 4(a) shows an example of aerosol transport from cropfires in Northern India to Jaipur on the 4th day Pre-Diwali in 2020, where the Corrected Reflectance 342 343 (True Color) and Deep Blue Aerosol Optical Thickness data taken by the VIIRS sensor (Fig. 4(a)) shows very high AOD values and thick smoke over Jaipur. Similarly, the surface measured PM_{2.5} 344 345 (Fig. 3(a)) and MODIS derived AOD (Fig. 2(c)) also showed a significant spike in values on the 4th day Pre-Diwali. Whereas the VIIRS sensor data on the 13th day Post-Diwali (Fig. 4(b)) shows 346 347 very low AOD distribution and a clear sky over Jaipur, similarly the lowest surface measured PM_{2.5} (Fig. 3(a)) was also observed on the 13th day Post-Diwali during the Diwali week. 348

349

350 **3.4** Air pollution observed by mobile monitoring during Diwali 2020

Reference grade monitors have high temporal coverage, but due to their static nature, they fail to provide information about the spatiotemporal air pollution variation in a region. Mobile monitoring with low-cost air quality sensors gives us the ability to have a wide spatial coverage with the same sensor unit and generate spatiotemporal data at fraction of cost. In the past, there have been few studies focused on mobile air quality monitoring in India by integrating low-cost air quality sensors on a car. <u>Agarwal et al., 2020</u> used an optical $PM_{2.5}$ sensor and retrofitted around 20 OLA cabs with them to perform mobile air quality monitoring in Delhi. <u>CSTEP and ILK Labs,</u> <u>2020</u> used DustTrak II aerosol monitor to measure $PM_{2.5}$, MicroAeth to measure Black Carbon and condensation particle counter to measure Ultrafine particles, and performed monitoring for 110 days using a CNG car in Bengaluru.

Fig. 5 shows the 3 different paths (Path A, B, C) we chose for mobile monitoring in Jaipur and the spatial $PM_{2.5}$ concentrations observed on different paths during different days of the experiment. We performed the mobile monitoring experiment on Path A on 11th November between 8:39 am to 9:06 am and observed very high mean $PM_{2.5}$ concentration (238.8 µg/m³) and spatial variability (SD = 53.6).

We performed the mobile monitoring experiment on Path B 5 times between 13th - 15th 366 November. On 13th November, we performed the experiment during morning and night time. 367 During the morning experiment on Path B1 between 9:15 am to 10:10 am, we observed both high 368 369 mean PM_{2.5} concentration (115.3 μ g/m³) and spatial variability (SD = 31.2). During the night 370 experiment on Path B2 between 9:02 pm to 9:50 pm, the mean PM_{2.5} concentration increased to 371 (196.5 μ g/m³) and showed high spatial variability (SD = 34.2). This increase could be due to the 372 dense traffic at night time, as lots of people go for shopping activities a day prior to Diwali. On 14th 373 November (Diwali day), we performed the experiment on Path B3 between 4:07 pm to 4:41 pm 374 and observed a drop in both mean PM_{2.5} concentrations (93.4 μ g/m³) and spatial variability (SD = 15.6). Similar results were observed during the experiment on Diwali night, on 15th November 375 376 between 12:02 am to 12:37 am. The mean PM_{2.5} concentration observed on Path B4 during the 377 mobile monitoring experiment on Diwali night was (92.3 μ g/m³), with very little spatial variability (SD = 9.6). Due to high wind speeds and the firecracker ban imposed during Diwali 2020, no spike 378

in PM_{2.5} concentration was observed from mobile monitoring on Diwali night. On 15th November, the morning following Diwali, we performed the experiment on Path B5 between 11:30 am to 12:30 pm, and observed a rise in both mean PM_{2.5} concentration (123 μ g/m³) and spatial variability (SD = 19.7).

On 16th November, we performed the mobile monitoring experiment on Path-C between 6:45 am to 9:40 am. Due to heavy rains and high wind speeds (SF-3, SF-6) on 15th November, we observed a drop in the mean $PM_{2.5}$ concentration (92.2 µg/m³) on 16th November with high spatial variability (SD = 31.5).

387 Table 2 shows the path travelled, mean $PM_{2.5} \pm SD$ value, distance covered (km), and duration of monitoring (minutes) for each mobile monitoring experiment. From Table 2, we can observe 388 389 that PM_{2.5} levels were significantly higher Pre-Diwali than during and Post-Diwali. During Diwali day, the PM_{2.5} levels were almost consistent, with mean PM_{2.5} being 93.4 ± 15.6 and 92.3 ± 9.6 390 391 $\mu g/m^3$ in the evening and night experiment respectively. Similar to the observations from the 392 reference-grade monitor during Diwali 2020 (Fig. 2(a)), we did not observe any spike in PM_{2.5} 393 concentrations from mobile monitoring on Diwali night. The total distance travelled on different 394 paths in Jaipur city during mobile monitoring was 108.04 km and the total duration of sensing was 395 265 minutes.

Fig. 6 shows the time-series plot of $PM_{2.5}$ concentrations observed during mobile monitoring. The purple air sensor has a temporal resolution of 1 second. From Fig. 6, we can observe the $PM_{2.5}$ variation on each path with time. On Path A, we observed high spatiotemporal variability, as $PM_{2.5}$ concentrations changed significantly with time. Similarly, high spatiotemporal variability could be observed on Paths B1, B2, and C. Whereas on Paths B3, B4, and B5, $PM_{2.5}$ showed less variability with time. 402 From the mobile monitoring plots on different paths (Fig. 5), we located a pollution hotspot. Fig. S9 shows the location of the pollution hotspot, where we observed a spike in PM_{2.5} 403 404 concentrations during the mobile monitoring experiment. Due to the ongoing flyover construction, 405 this region is packed with traffic jams and the construction activity leads to an increase in the 406 concentration of dust particles. The data we generated from our experiment shows how impactful 407 mobile monitoring could be to generate fine-grained PM2.5 data for regions like Tier-2 cities and 408 rural areas, that currently lack sufficient monitoring capabilities. In 2017, a study found that there 409 should be around 26 reference-grade monitors in Jaipur to efficiently monitor pollution levels in 410 the city (Urban Emissions, 2017), but the number currently stands at only 3. Thus, the accurate 411 results we achieved from our experiment clearly show that low-cost mobile air quality monitoring 412 can bridge the gap in the lack of monitoring capabilities for Indian cities at a fraction of cost.

413

414 4 DISCUSSIONS

415

The rising impact of urbanisation, post-monsoon stubble burning from northern states of India, and dust particles from the Thar desert have contributed to the rise in AOD over Jaipur from 2000-2020. While the $PM_{2.5}$ concentrations were found to be substantially higher than the annual mean safe limit of $10\mu g/m^3$ for the past 4 years, showing the impact of air pollution on even the Tier-2 cities of India.

From the temporal analysis of $PM_{2.5}$ during the Diwali festival, firecracker activities were found to be the primary cause of the pattern of a sudden rise in hourly $PM_{2.5}$ concentrations, which peaked at the night-time during Diwali days in the previous years (2017-19). The ban imposed by the State Government on bursting firecrackers due to Covid-19 during Diwali in 2020, prevented this pattern of sudden elevation in hourly $PM_{2.5}$ concentrations from occurring during the night-time of Diwali, showing the impact of firecrackers on air quality. A significant impact of wind speed and RH was also observed on $PM_{2.5}$ concentrations during Diwali. The increasing trend in wind speed and RH from Pre to Post Diwali days in 2020, led to a drop in daily mean $PM_{2.5}$ concentrations from Pre to Post Diwali days, whereas the opposite trend in $PM_{2.5}$ concentrations was observed in the previous years (2017-19).

From the spatiotemporal analysis of PM_{2.5} during the Diwali festival on different days and paths, mobile monitoring was found to be a cheap substitute for static reference-grade monitors. Mobile monitoring using low-cost sensor showed patterns similar to reference-grade monitors, showing its potential as a cheap alternate for performing air quality in Tier-2 cities of India.

- 435
- . . .

436 5 SUMMARY AND CONLUSIONS

437

Diwali festival creates a unique environmental condition where emissions of air pollutants from 438 439 different sources like firecrackers, increased human (and industrial) activities and crop-fires get 440 combined and drastically deteriorates the air quality of IGP and neighbouring regions. Calm winds 441 and low temperatures coincident with the festival season favour the poor air quality conditions. To 442 curb the COVID-19 pandemic in Jaipur, the government had imposed a ban on bursting firecrackers 443 during Diwali 2020. In this study, to evaluate the spatiotemporal variability in particulate matter 444 (or aerosols) during the Diwali festival, we performed a comprehensive analysis using data from 445 the MODIS and VIIRS sensor, three regulatory ground monitors, and mobile monitoring using a 446 low-cost sensor. The end goal of the analysis is to assess the impact of the firecracker ban and 447 meteorological aspects on air pollution during Diwali. In addition, we have also analysed long-448 term trends in particle pollution over the city.

450	The key findings from our analysis are:
451	• The climatological AOD derived from the two MODIS (Aqua and Terra) sensors over Jaipur
452	and Delhi in the period 2000-2020 was 0.3 ± 0.04 and 0.72 ± 0.06 respectively with a 29%
453	increase over the period.
454	• The mean (2017-2020) $PM_{2.5}$ concentrations in Jaipur. (67 ± 34.6 µg/m ³) is significantly lower
455	compared to the capital city of New Delhi (102.6 \pm 93.5 μ g/m ³).
456	• The firecracker ban in Jaipur during 2020 Diwali was effective in keeping the pollution levels
457	in check whereas a significantly elevated level of pollution was found during the previous
458	Diwali festivals.
459	• The three regulatory monitors in the city consistently measured similar $PM_{2.5}$ values during
460	Diwali, demonstrating very little spatial variability.
461	• Smoke aerosol transport from crop-fires in Northern India to Jaipur was observed in satellite
462	data and elevated both column (i.e. AOD) and surface ($PM_{2.5}$) pollution levels on 10^{th}
463	November, 2020. The AOD increased by 813 % and $PM_{2.5}$ increased by 111% between 6 th
464	November and 10 th November, 2020.
465	• Mobile monitoring in different parts of the city also did not show any spike in $PM_{2.5}$
466	concentrations during Diwali.
467	• Mobile monitoring was able to identify a pollution hotspot near an under-construction flyover
468	in the city.
469	
470	ACKNOWLEDGMENTS
471 472	Pawan Gupta was supported by the NASA ROSES program NNH17ZDA001N-TASNPP: The
473	Science of Terra, Aqua, and Suomi NPP. MODIS and VIIRS data were obtained from NASA

474 LAADS and Worldview. The ground-based air quality monitoring data was obtained through the

475	data portal of the Central Pollution Control Board (CPCB), New Delhi, India. Piyush Yadav would
476	like to thank Prof. S. Indu, DTU, India for providing technical guidance about sensors.
477	
478	REFERENCES
479	A special report on global exposure to air pollution and its disease burden, State Of Global Air,
480	2019. Website: https://www.stateofglobalair.org/sites/default/files/soga_2019_report.pdf
481	A special report on global exposure to air pollution and its health impacts, State Of Global Air,
482	2020. Website: <u>https://www.stateofglobalair.org/sites/default/files/documents/2020-10/soga-</u>
483	<u>2020-report-10-26_0.pdf</u>
484	Air Quality Life Index (2021). Annual
485	Update. Website:
486	https://aqli.epic.uchicago.edu/wp- content/uploads/2021/08/AQLI_2021_Report
487	English.IndiaVersion.pdf.
488	(Accessed: October, 2021).

- 489 Agarwal, D.; Iyengar, S.; Swaminathan, M.; Sharma, E.; Raj, A.; Hatwar, A. Modulo: Drive-by
- 490 Sensing at City- scale on the Cheap. In ACM SIGCAS Conference on Computing and Sustain- able
- 491 Societies (COMPASS '20), June 15–17, 2020, Ecuador. ACM, New York, NY, USA, 11 pages.
- 492 <u>https://doi.org/10.1145/3378393.3402275</u>

493 Attri AK, Kumar U, Jain VK. 2001. Formation of ozone by fireworks. Nature 411:1015. DOI:
494 10.1038/35082634

495 Census India, Website: <u>https://censusindia.gov.in/2011-prov-</u>
496 <u>results/paper2/data_files/India2/Table_2_PR_Cities_1Lakh_and_Above.pdf</u>. (Accessed: January,
497 2021).

- Chattopadhyay. V, Pollution no less this Diwali, shows official data, Centre for Science and
 Environment, Oct. 24, 2014. Website: <u>https://www.downtoearth.org.in/news/pollution-no-less-this-</u>
 <u>diwali-shows-official-data-47063 (Accessed: December, 2020).</u>
- Chen, L.; Zhu, J.; Liao, H.; Yang, Y.; Yue, X. Meteorological influences on PM2.5 and O3
 trends and associated health burden since China's clean air actions, Science of The Total
 Environment, Volume 744, 2020, 140837, ISSN 0048-9697,
 <u>https://doi.org/10.1016/j.scitotenv.2020.140837</u>.
- Chowdhury, S.; Dey, S.; Di Girolamo, L.; Smith, K.R.; Pillarisetti, A.; Lyapustin, A. Tracking
 ambient PM_{2.5} buildup in Delhi national capital region during the dry season over 15 years using
 a high-resolution (1-km) satellite aerosol dataset. *Atmos. Environ.* 2019, *204*, 142–150. [CrossRef]
- 508 Christopher, S.; Gupta, P. Global distribution of column satellite aerosol optical depth to surface
 509 PM2.5 relationships Remote Sensing, 12 (2020), p. 1985, 10.3390/rs12121985

510	Cichowicz, R., Wielgosiński, G. & Fetter, W. Effect of wind speed on the level of particulate
511	matter PM10 concentration in atmospheric air during winter season in vicinity of large combustion
512	plant. J Atmos Chem 77, 35–48 (2020). <u>https://doi.org/10.1007/s10874-020-09401-w</u>
513	CPCB. Website: <u>https://app.cpcbccr.com</u> . (Accessed: December, 2020).
514	CPCB. Air quality monitoring, emission inventory and source apportionment study for Indian
515	cities: National Summary Report, Central Pollution Control Board, New Delhi, India (2011)
516	CSTEP & ILK Labs (2020). Mobile-monitoring campaign for air pollution studies in Bengaluru.
517	(CSTEP-WS-2020-04). Website: <u>https://cstep.in/drupal/node/1296</u> . (Accessed: December, 2020).
518	Cusworth, D.H.; Mickley, L.J.; Sulprizio, M.P.; Liu, T.; Marlier, M.E.; DeFries, R.S.;
519	Guttikunda, S.K.; Gupta, P. Quantifying the influence of agricultural fires in northwest India on
520	urban air pollution in Delhi, India. Environ. Res. Lett. 2018, 13, 044018.
521	DeepBlue, NASA. Website: https://deepblue.gsfc.nasa.gov/science. (Accessed: January, 2021).
522	Deka P, Hoque RR. 2014. Incremental effect of festive biomass burning on wintertime PM_{10} in
523	Brahmaputra Valley of Northeast India. Atmospheric Research 143:380-391. DOI:
524	10.1016/j.atmosres.2014.03.003

525	Devi, N.L.; Kumar, A.; Yadav, I.C. PM10 and PM2.5 in Indo-Gangetic Plain (IGP) of India:
526	Chemical characterization, source analysis, and transport pathways, Urban Climate, Volume
527	33,2020,100663,ISSN 2212-0955, https://doi.org/10.1016/j.uclim.2020.100663.
528	Dey, S.; Purohit, B.; Balyan, P.; Dixit, K.; Bali, K.; Kumar, A.; Imam, F.; Chowdhury, S.; Ganguly,

- 529 D.; Gargava, P.; Shukla, V.K. A Satellite-Based High-Resolution (1-km) Ambient PM_{2.5} Database
- 530 for India over Two Decades (2000–2019): Applications for Air Quality Management. Remote
- 531 Sens. 2020, 12, 3872. https://doi.org/10.3390/rs12233872
- 532 DEY, Sharadia et al. Influences of boundary layer phenomena and meteorology on ambient air

quality status of an urban area in eastern India. Atmósfera, [S.l.], v. 31, n. 1, p. 69-86, dec. 2017.

534 ISSN 2395-8812. Available at:

535 <u>https://www.revistascca.unam.mx/atm/index.php/atm/article/view/ATM.2018.31.01.05</u>. Date

536 accessed: 22 jan. 2021. doi:<u>https://doi.org/10.20937/ATM.2018.31.01.05</u>.

Donkelaar, A.V.; Martin, R.V.; Li, C.; Burnett, R.T. Regional Estimates of Chemical 537 Composition of Fine Particulate Matter Using a Combined Geoscience-Statistical Method with 538 539 Information from Satellites, Models, Environmental æ and Monitors. Science Technology 2019 53 (5), 2595-2611, DOI: https://doi.org/10.1021/acs.est.8b06392 540

541	Environmental	Risk	Outlook,	2021.
542	Website: <u>https://www.maplec</u>	croft.com/insights/analy	vsis/environmental-risk-outlook	<u>-</u>
543	2021/#report form containe	r		

544	Flower, V.J.B.; Kahn, R.A. Interpreting the Volcanological Processes of Kamchatka, Based on
545	Multi-Sensor Satellite Observations. Remote Sens. Environ. 2020, 237, 111585.

546 Ganguly ND. 2009. Surface ozone pollution during the festival of Diwali, New Delhi, India547 Earth Science India 2:224-229.

Ganguly ND. 2015. Short term change in relative humidity during the festival of Diwali in India.
Journal of Atmospheric and Solar-Terrestrial Physics 129:49-54. DOI: 10.1016/j.jastp.2015.04.007

550 Ganguly, N., C. Tzanis, K. Philippopoulos and D. Deligiorgi. Analysis of a severe air pollution

episode in India during Diwali festival – a nationwide approach., *Atmosfera* 32 (2019): 225-236.

552 Gupta, P.; Christopher, S.A.; Box, M.A.; Box, G.P. Multi year satellite remote sensing of 553 particulate matter air quality over Sydney, Australia. Int. J. Remote Sens. 2007, 28, 4483–4498.

Gupta, P.; Christopher, S.A.; Wang, J.; Gehrig, R.; Leed, Y.; Kumar, N. Satellite remote sensing
of particulate matter and air quality assessment over global cities, Atmospheric Environment, 40
(2006), pp. 5880-5892, ISSN 1352-2310, <u>https://doi.org/10.1016/j.atmosenv.2006.03.016</u>

Guttikunda, S.K.; Goel, R.; Pant, P. Nature of air pollution, emission sources, and management
 in the Indian cities, Atmospheric Environment, Volume 95,2014, Pages 501-510, ISSN 1352-2310,
 <u>https://doi.org/10.1016/j.atmosenv.2014.07.006</u>

560	Leung,	F.	2021.	How	China	is	Winning	Its	Battle	Against	Air	Pollution.	Website	:
							-			-				
561	https://ear	th.o	rg/how-	-china-	is-winn	ing	-its-battle-	agai	nst-air-j	pollution/				

Huff, A.K.; Kondragunta, S. Meteorologists track wildfires using satellite smoke images. Eos
2017, 98, 18–23.

India – Ambient Air Monitoring Data. Website: <u>https://urbanemissions.info/india-air-</u>
 <u>quality/india-ambient-monitoring-data/</u>

Jayamurugan, R.; Kumaravel, B.; Palanivelraja, S.; Chockalingam, M. P. Influence of
Temperature, Relative Humidity and Seasonal Variability on Ambient Air Quality in a Coastal
Urban Area, International Journal of Atmospheric Sciences, vol. 2013, Article ID 264046, 7 pages,
2013. https://doi.org/10.1155/2013/264046

Kulshrestha UC, Nageswara RT, Azhaguvel S Kulshrestha MJ. 2004. Emissions and
accumulation of metals in the atmosphere due to crackers and sparkles during Diwali festival in
India. Atmospheric Environment 38:4421-4425. DOI: 10.1016/j.atmosenv.2004.05.044

Kumar, M.; Parmar, K.S.; Kumar, D.B.; Mhawish, A.; Broday, D.M.; Mall, R.K.; Banerjee, T.
Long-term aerosol climatology over Indo-Gangetic Plain: Trend, prediction and potential source
fields, Atmospheric Environment, Volume 180, 2018, Pages 37-50, ISSN 1352-2310,
<u>https://doi.org/10.1016/j.atmosenv.2018.02.027</u>.

Liu, Y.; Paciorek, C.J.; Koutrakis, P. Estimating regional spatial and temporal variability of
PM(2.5) concentrations using satellite data, meteorology, and land use information. *Environ Health Perspect*. 2009;117(6):886-892. doi: 10.1289/ehp.0800123

Liu, Y., Zhou, Y. & Lu, J. Exploring the relationship between air pollution and meteorological
conditions in China under environmental governance. Sci Rep 10, 14518 (2020).
https://doi.org/10.1038/s41598-020-71338-7

Mangla, R.; J, I.; S.S., C. Inter-comparison of multi-satellites and Aeronet AOD over Indian
Region,Atmospheric Research,Volume 240,2020,104950,ISSN 0169-8095,
<u>https://doi.org/10.1016/j.atmosres.2020.104950</u>

Mittal SK, Singh N, Agarwal R, Awasthi A Gupta PK. 2009. Ambient air quality during wheat
and rice crop stubble burning episodes in Patiala. Atmospheric Environment 43:238-244. DOI:
10.1016/j.atmosenv.2008.09.068

Naeger, A.R.; Gupta, P.; Zavodsky, B.T.; McGrath, K.M. Monitoring and tracking the transPacific transport of aerosols using multi-satellite aerosol optical depth composites. Atmos. Meas.
Tech. 2016, 9, 2463–2482.

Ojha, N., Sharma, A., Kumar, M. et al. On the widespread enhancement in fine particulate matter
across the Indo-Gangetic Plain towards winter. Sci Rep 10, 5862 (2020).
<u>https://doi.org/10.1038/s41598-020-62710-8</u>

595	Pandey, A.; Brauer, M.; Cropper, M.L.; Balakrishnan, K.; Mathur, P.; Dey, S.; Turkgulu, B.;
596	Kumar, G.A.; Khare, M.; Beig, G.; et al. Health and economic impact of air pollution in the states
597	of India: The Global Burden of Disease Study 2019. Lancet Planet. Health 2020. [CrossRef]
598	PIB India, India achieves 100 GW Milestone of Installed Renewable Energy Capacity. Website:
599	https://pib.gov.in/PressReleasePage.aspx?PRID=1745254
600	Pozzer, A.; Dominici, F.; Haines, A.; Witt, C.; Münzel, T.; Lelieveld, J.; Regional and global
601	contributions of air pollution to risk of death from COVID-19, Cardiovascular Research, Volume
602	116, Issue 14, 1 December 2020, Pages 2247–2253, <u>https://doi.org/10.1093/cvr/cvaa288</u>
603	PurpleAir. Website: <u>https://www2.purpleair.com/pages/technology</u>
604	Ramachandran, S.; Kedia, S.; Srivastava, R. Aerosol optical depth trends over different regions
605	of India, Atmospheric Environment, Volume 49, 2012, Pages 338-347, ISSN 1352-2310,
606	https://doi.org/10.1016/j.atmosenv.2011.11.017.
607	Saha U, Talukdar S, Jana S, Maitra A. 2014. Effects of air pollution on meteorological
608	parameters during Deepawali festival over an Indian urban metropolis. Atmospheric Environment
609	98:530-539.

610 DOI: 10.1016/j.atmosenv.2014.09.032

611 Sarkawt M.L. Hama, Prashant Kumar, Roy M. Harrison, William J. Bloss, Mukesh Khare, Sumit Mishra, Anil Namdeo, Ranjeet Sokhi, Paul Goodman, Chhemendra Sharma, Four-year assessment 612 613 of ambient particulate matter and trace gases in the Delhi-NCR region of India, Sustainable Cities 614 Volume 2020, 102003, ISSN and Society, 54, 2210-6707, 615 https://doi.org/10.1016/j.scs.2019.102003.

Sathe, Y.; Kulkarni, S.; Gupta, P.; Kaginalkar, A.; Islam, S.; Gargava, P. Application of
Moderate Resolution Imaging Spectroradiometer (MODIS) Aerosol Optical Depth (AOD) and
Weather Research Forecasting (WRF) model meteorological data for assessment of fine particulate
matter (PM2.5) over India Atmospheric Pollution Research, 10 (2) (2019), pp. 418434, 10.1016/j.apr.2018.08.016

Sindhwani, R.; Goyal, P. Assessment of traffic-generated gaseous and particulate matter
emissions and trends over Delhi (2000–2010), Atmospheric Pollution Research, Volume 5, Issue
3,2014, Pages 438-446, ISSN 1309-1042, <u>https://doi.org/10.5094/APR.2014.051</u>.

Thaker P, Gokhale S. The impact of traffic-flow patterns on air quality in urban street canyons.
Environ Pollut. 2016 Jan;208(Pt A):161-169. doi: 10.1016/j.envpol.2015.09.004. Epub 2015 Sep
26. PMID: 26412198.

- 627 The Air Pollution Knowledge Assessment (APnA) City, Urban Emissions, 2017. Website:
- 628 <u>https://urbanemissions.info/wp-content/uploads/apna/docs/india_apna_2017_jaipur.pdf</u>
- 629 (Accessed: December, 2020).

The Times of India, Website: https://timesofindia.indiatimes.com/city/jaipur/raj-govt-bans-

631	firecrackersafter-assessing-covid-risk/articleshow/78989259.cms
632	Twitter, Website: https://twitter.com/ashokgehlot51/status/1323130038867259394?s=20
633 634 635	Verma, S., Payra, S., Gautam, R. <i>et al.</i> Dust events and their influence on aerosol optical properties over Jaipur in Northwestern India. <i>Environ Monit Assess</i> 185, 7327–7342 (2013). https://doi.org/10.1007/s10661-013-3103-9
636	Wang, J.; Christopher, S.A. Christopher Mesoscale modeling of Central American smoke
637	transport to the United States: 2. Smoke radiative impact on regional surface energy budget and
638	boundary layer evolution. J. Geophys. Res. 2006, 111, D14S92.
639	Wang Q., Urbanization and Global Health: The Role of Air Pollution, Iranian journal of public
640	health,2018, 47(11), 1644–1652.
641	WHO Air Pollution. Website: <u>https://www.who.int/health-topics/air-pollution#tab=tab_1</u>
642	WHO Global Urban Ambient Air Pollution Database (update 2016). Website:
643	https://www.who.int/phe/health_topics/outdoorair/databases/cities/en/#:~:text=According%20to
644	%20the%20latest%20urban,that%20percentage%20decreases%20to%2056%25
645	Yadav, P., Purple-Air-Live-Data, Github. Website: <u>https://github.com/piyushy6/Purple-Air-</u>
646	Live-Data. (Accessed: December, 2020).
	29

647 Yadav, P., Rpi-Pixhawk-GPS, Github. Website: <u>https://github.com/piyushy6/Rpi-Pixhawk-</u>
648 GPS. (Accessed: December, 2020).

649 Yang, Q.; Yuan, Q.; Yue, L.; Li, T.; Shen, H.; Zhang, L. The relationships between PM2.5 and

aerosol optical depth (AOD) in mainland China: About and behind the spatio-temporal variations,

651 Environmental Pollution, 248 (2019), pp. 526-535, doi: <u>10.1016/j.envpol.2019.02.071</u>

652 Yu, H.; Tan, Q.; Chin, M.; Remer, L.A.; Kahn, R.A.; Bian, H.; Kim, D.; Zhang, Z.; Yuan, T.;

653 Omar, A.; et al. Estimates of African Dust Deposition Along the Trans-Atlantic Transit Using the

654 Decadelong Record of Aerosol Measurements from CALIOP, MODIS, MISR, and IASI. J.

655 Geophys. Res. Atmos. 2019, 124, 7975–7996..

656

657

658

659

660

661

662

664	
665	Table 1. Mean $PM_{2.5} \pm \sigma (\mu g/m^3)$ value of 3-Days Pre and Post Diwali for years 2017-20.
666	
667	

Diwali	Pre-Diwali Days			Diwali Day	Post-Diwali Days			
Week Days	-3	-2	-1		+1	+2	+3	
Year								
2017	140 ± 86	145 ± 59	158 ± 41	245 ± 231	172 ± 188	195 ± 173	131 ± 66	
2018	113 ± 59	63 ± 34	114 ± 49	114 ± 136	136 ± 130	171 ± 109	166 ± 127	
2019	77 ± 28	110 ± 26	83 ± 30	128 ± 174	226 ± 224	222 ± 154	80 ± 32	
2020	113 ± 36	88 ± 33	110 ± 36	106 ± 37	79 ± 15	63 ± 15	56 ± 18	
669 670								
671								
672								

678	Table 2. Mean $PM_{25} \pm \sigma$	(ug/m3) value observ	ed by performing r	mobile monitoring on 3 different
0,0	Habit 2. Inform $1 \operatorname{Inf}_{2,3} = 0$	(mg/ille) fulue obbert	ca by perioriting i	moone monitoring on 5 anierent

679 Paths (A, B, C) of Jaipur during 2020 Diwali week.

680 681

682							
Date & Time	11/11 (8:39am -	13/11 (9:15am -	13/11 (9:02pm -	14/11 (4:07pm -	15/11 (12:02am -	15/11 (11:30am -	16/11 (6:45am -
	9:06am)	10:10am)	9:50pm)	4:41pm)	12:37am)	12:30pm)	9:40am)
Path Travelled	A	B1	B2	B3	B4	B5	C
Mean $PM_{2.5} \pm \sigma$	238.8 ± 53.6	115.3 ± 31.2	196.5 ± 34.4	93.4 ± 15.6	92.3 ± 9.6	123 ± 19.7	92.2 ± 31.5
Distance Covered (km)	8.63	11.97	13.14	12	13.9	15.3	33.1
Duration of Monitoring (Minutes)	27	33	48	34	27	37	59
Reference Monitor 24-hr mean PM _{2.5}	113 ± 36	110 ± 36	-	106 ± 37	-	79 ± 15	63 ± 15
				-			

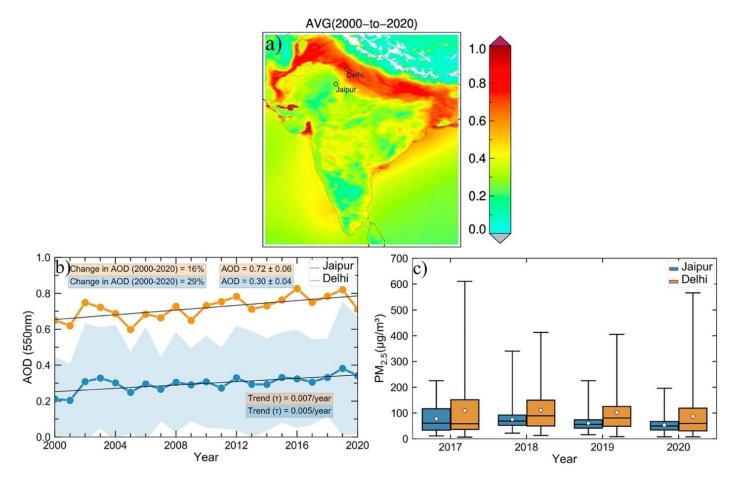


Fig. 1. a) Mean aerosol optical depth (2000-2020) derived from two MODIS sensors over India.
The marked is location of Delhi and Jaipur, b) Trends in AODs over the same period for two cities
extracted using same data sets; shaded area shows one standard deviation over Jaipur, c) Annual
mean PM_{2.5} trends (2017-2020) in two cities (Station name – Jaipur : Police Commissionerate ;
Delhi : Mandir Marg).

689

690

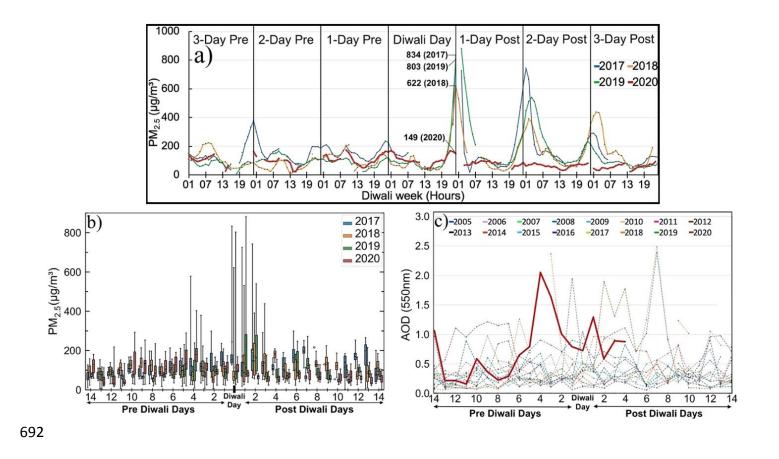


Fig. 2. a) Reference Monitor 1-hour mean PM_{2.5} data of 3-Days Pre and Post Diwali for years 201720, hours are in local Indian standard time, b) Reference Monitor Box-Plot using 1-hour mean
PM_{2.5} data of 14-Days Pre and Post Diwali for years 2017-20, c) MODIS AOD data of 14-Days
Pre and Post Diwali for years 2005-2020.

697

698

699

700

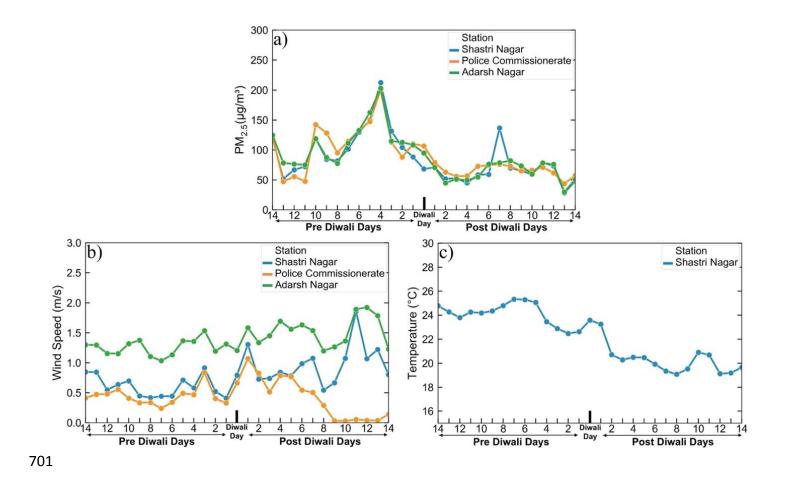
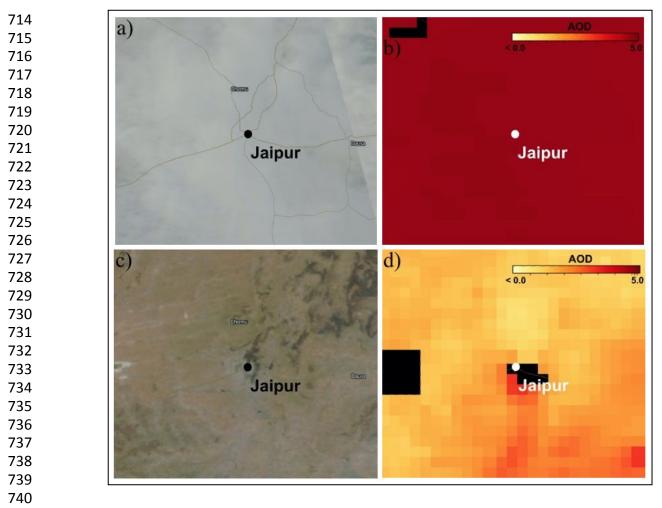


Fig. 3. Daily mean evolution of a) PM_{2.5}, b) Wind Speed, c) Temperature observed in Jaipur's 3-

703 Reference Monitor for +-14 days of 2020 Diwali Week.



741

742 743 Fig. 4. Impact of aerosol transport over Jaipur during 2020 Diwali week observed from VIIRS 744 745 sensor. a) shows Corrected Reflectance (True Color) and b) shows Deep Blue Aerosol Optical Thickness over Jaipur on 10/11/2020 (4th day Pre-Diwali) ; c) shows Corrected Reflectance (True 746 Color) and d) shows Deep Blue Aerosol Optical Thickness over Jaipur on 27/11/2020 (13th day 747

- 748 Post-Diwali).
- 749
- 750
- 751

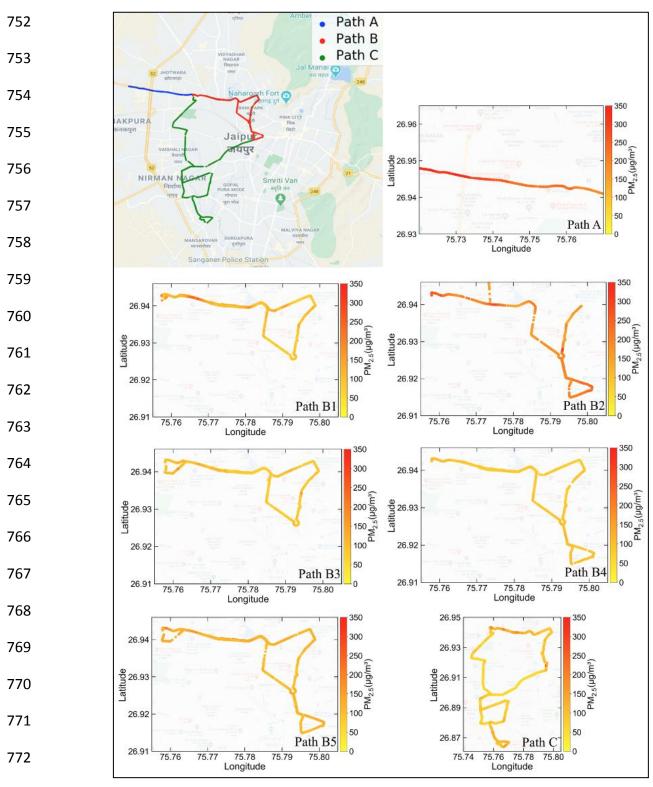
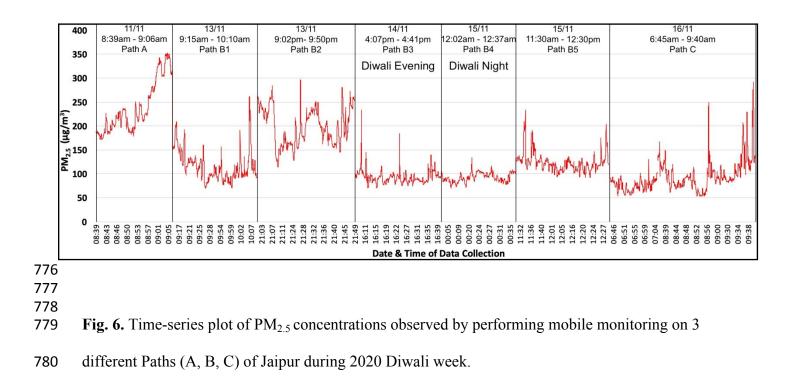


Fig. 5. Mobile air quality monitoring performed on 3 different Paths (A, B, C) of Jaipur by
integrating air quality sensor on the roof of a car. Data collected from - Path A on 11/11; Path B1
and Path B2 on 13/11; Path B3 and Path B4 on 14/11; Path B5 on 15/11; Path C on 16/11.



Supplement Material

Paper Title: Temporal and Spatial Variability in Particulate Matter Pollution in The Pink City during the 2020 Diwali Festival

Supplement Table

Year	Diwali Date		
2005	01-Nov-05		
2006	21-Oct-06		
2007	09-Nov-07		
2008	28-Oct-08		
2009	17-Oct-09		
2010	05-Nov-10		
2011	26-Oct-11		
2012	13-Nov-12		
2013	03-Nov-13		
2014	23-Oct-14		
2015	11-Nov-15		
2016	30-Oct-16		
2017	19-Oct-17		
2018	07-Nov-18		
2019	27-Oct-19		
2020	14-Nov-20		

Table S1. Diwali dates for the past 15 years

	Pre-Diwali Days		Diwali Day	Post-Diwali Days	
Diwali Week Days					
	-2	-1		+1	+2
Year					
2017	210	237	834	743	291
2018	145	156	622	391	438
2019	114	121	881	540	232
2020	143	168	149	87	82

Table S2. Peak hourly PM2.5 (μ g/m³) values observed during the nighttime of 2-Days Pre and Post Diwali for years 2017-20

Supplement Figure

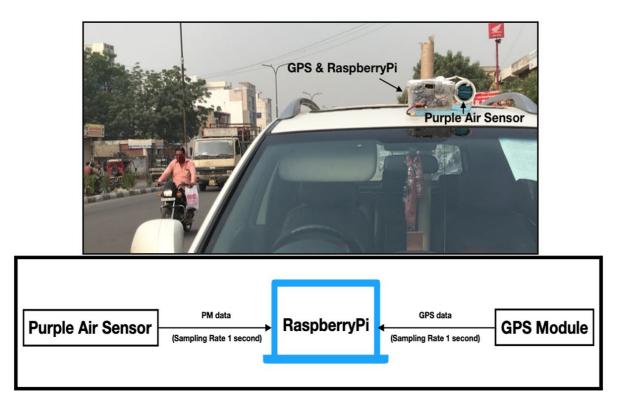


Fig. S1. Purple Air Sensor used for mobile monitoring placed on top of the car.

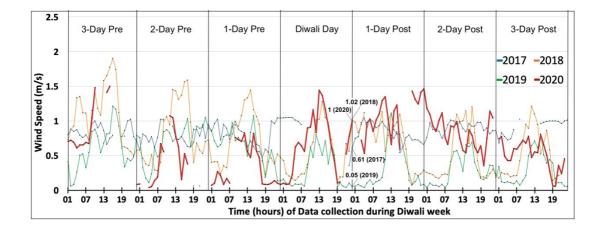


Fig. S2. Reference Monitor 1-hour mean wind speed (m/s) data of 3-Days Pre and Post Diwali for years 2017-20, hours are in local Indian standard time.

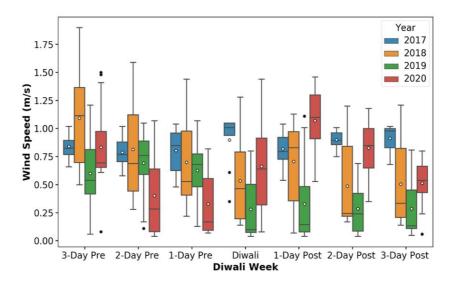


Fig. S3. Reference Monitor Box-Plot using 1-hour mean wind speed (m/s) data of 3-Days Pre and Post Diwali for years 2017-20.

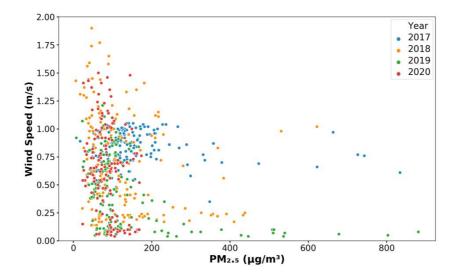


Fig. S4. Reference Monitor scatter plot using 1-hour mean wind speed (m/s) and $PM_{2.5}$ (µg/m³) data of 3-Days Pre and Post Diwali for years 2017-20.

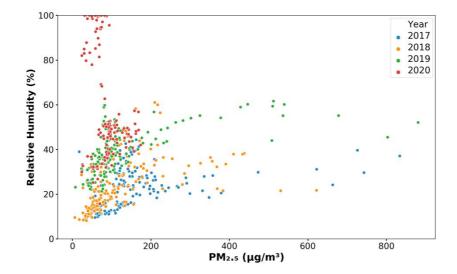


Fig. S5. Reference Monitor scatter plot using 1-hour mean relative humidity (%) and $PM_{2.5}$ (µg/m3) data of 3-Days Pre and Post Diwali for years 2017-20.

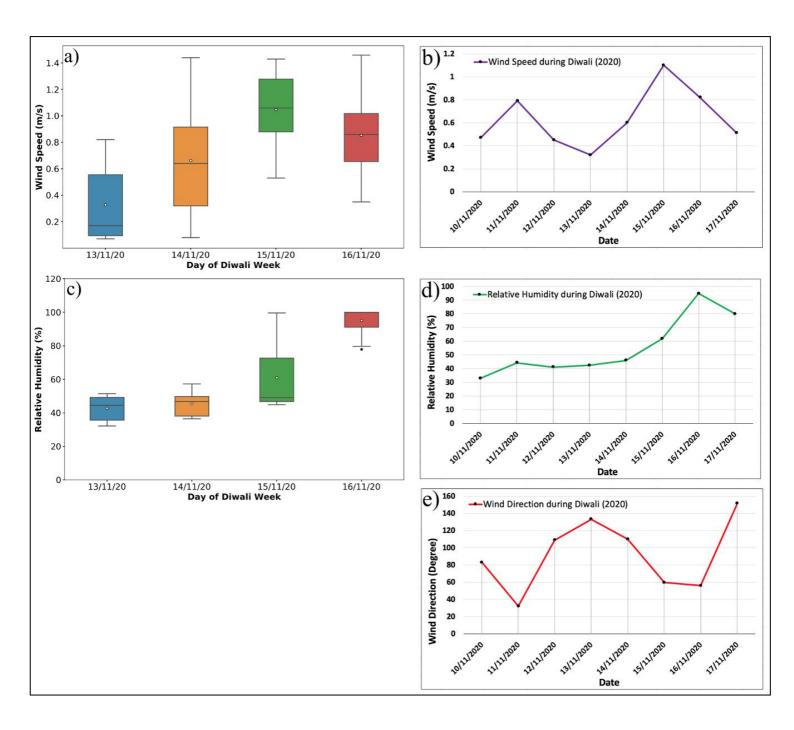


Fig. S6. a) Reference Monitor box plot using 1-hour mean wind speed (m/s) data during 4 days of 2020 Diwali week, b) Reference Monitor line plot using 1-hour mean wind speed (m/s) data during 8 days of 2020 Diwali week, c) Reference Monitor box plot using 1-hour mean relative humidity (%) data during 4 days of 2020 Diwali week, d) Reference Monitor line plot using 1-hour mean relative numidity (%) data during 8 days of 2020 Diwali week, e) Reference Monitor line plot using 1-hour mean relative humidity (%) data during 8 days of 2020 Diwali week, e) Reference Monitor line plot using 1-hour mean wind direction (degree) data during 8 days of 2020 Diwali week.



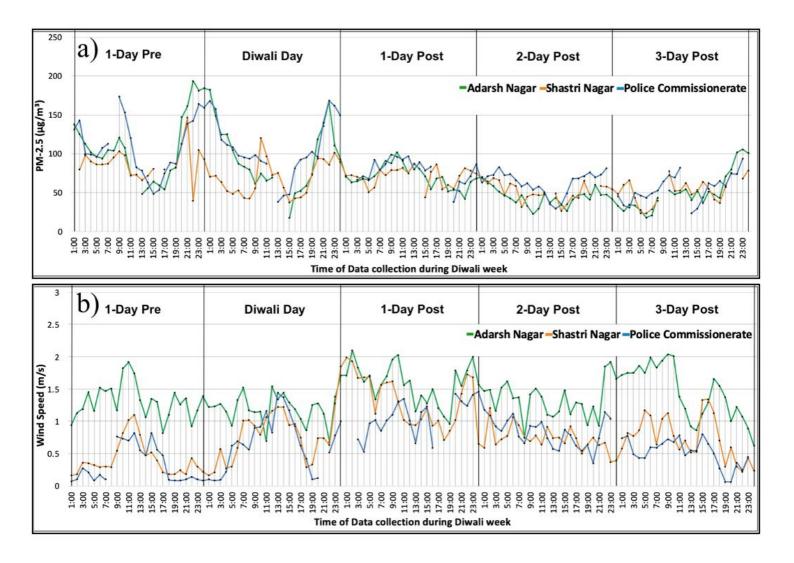


Fig. S7. 1-hour mean a) $PM_{2.5}$ (µg/m3), b) wind speed (m/s) data of the 5 day period during Diwali of the 3-different reference monitor of Jaipur, hours are in local Indian standard time.

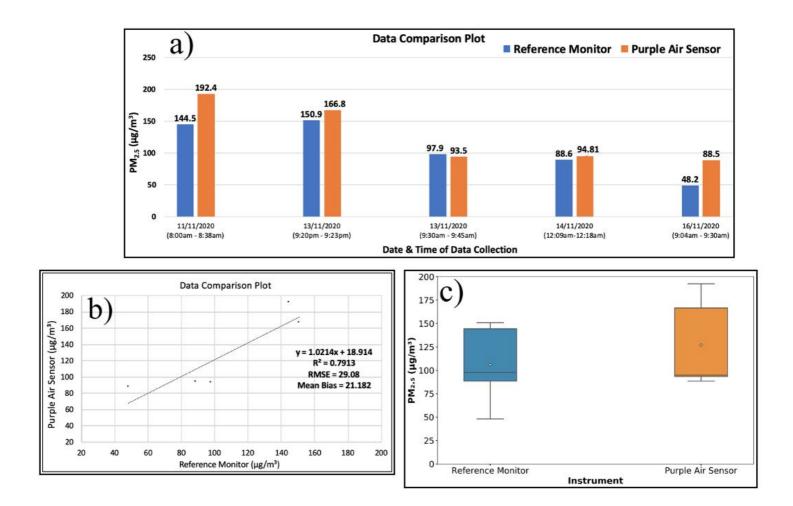


Fig. S8. Purple Air sensor data collocation results against a reference monitor (Shastri Nagar station) in Jaipur. a) Bar Plot, b) Line Plot, c) Box Plot

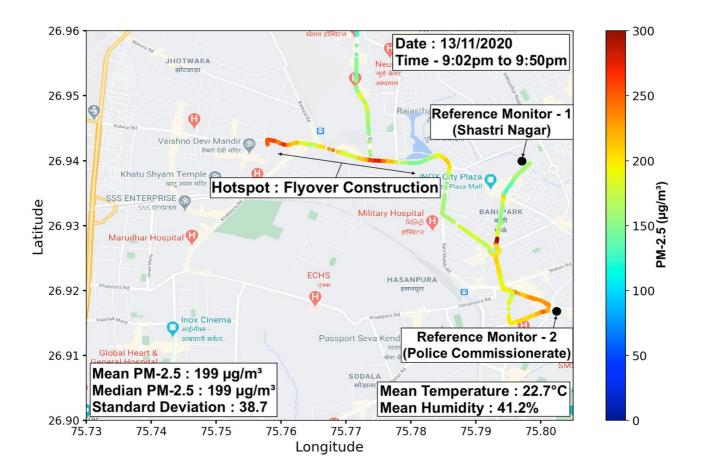


Fig. S9. Hotspot (construction site of a flyover) located by mobile monitoring experiment.